

A COMPARISON OF NATIVE VERSUS OLD-FIELD VEGETATION IN UPLAND PINELANDS MANAGED WITH FREQUENT FIRE, SOUTH GEORGIA, USA

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ABSTRACT

Fire-maintained, herb-dominated upland pinelands of the southeastern U.S. Coastal Plain may be broadly divided into those that have arisen through secondary succession following abandonment of agriculture (old-field pinelands) and those that have never been plowed (native pinelands). The ability to distinguish these habitat types is important for setting conservation priorities by identifying natural areas for conservation and appropriate management and for assessing the ecological value and restoration potential for old-field pine forests managed with frequent fire. However, differences in species composition have rarely been quantified. The goals of this study were to characterize the species composition of native and old-field pineland ground cover, test the ability to distinguish communities of previously unknown disturbance history, and suggest indicator species for native versus old-field pinelands. Plant composition was surveyed in areas known to be native ground cover, those known to be old fields, and those with an uncertain disturbance history. Twelve permanent plots were established in each cover type and sampled in spring (April–May) and fall (October–November) in 2004 and 2005. Of the 232 species identified in the plots, 56 species were present only in native ground-cover plots, of which 17 species occurred in a sufficient number of plots to have a statistically significant binomial probability of occurring in native ground cover and might be considered indicator species. In addition, 15 species were confined to old fields, of which 5 had a statistically significant binomial probability. Additionally, plant census transects from a previous survey were comparatively analyzed, yielding a total of 432 species, of which 111 were present only in native ground-cover transects and 3 occurred in a sufficient number of transects to have statistically significant binomial association with native areas. Also, 111 species were confined to old fields but none in a sufficient number of transects for a significant association with old fields. In both the plot and transect data sets, most old-field species represented a subset of those found in native areas, suggesting differential ability of certain native species to disperse to and become established in abandoned agricultural land. These results will assist in identifying natural areas for conservation as well as assessing the ecological value and restoration potential of old-field pine forests managed with frequent fire.

keywords: indicator species, longleaf pine, native ground cover, old fields, *Pinus palustris*, prescribed fire, secondary succession, South Georgia, upland pinelands.

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INTRODUCTION

A detailed knowledge of a region's natural communities is crucial for conservation and restoration efforts as well for the management of currently protected areas. This baseline knowledge allows for a determination of the degree to which a degraded system has changed and potentially concurrent loss of biodiversity and possible changes in ecosystem processes (Noss 2000). The ability to identify undisturbed natural communities may be obfuscated by a past history of anthropogenic disturbance or alteration of natural disturbance regimes. In the southeastern U.S. Coastal Plain, intensive agriculture is chief among anthropogenic activities that have contributed to a 97% conversion of native pineland habitat (Frost 1993). Thus, conservation efforts require the means to distinguish and characterize native versus post-agricultural communities in order to set conservation priorities for acquisition and

protection and to assess the ecological value and restoration potential of each.

Following the large-scale abandonment of agricultural fields on the Coastal Plain in the late 19th and early 20th centuries (Paisley 1968, Brueckheimer 1979), fallow fields that were periodically burned returned to pinelands with an herb-dominated understory (Moser et al. 2002). Thus, fire-maintained, herb-dominated upland pinelands of the southeastern U.S. Coastal Plain may be broadly divided into those that have arisen through secondary succession following abandonment of agriculture (old-field pinelands) and those that have never been plowed (native pinelands). With proactive management (i.e., prescribed burning and selective timber thinning), these old-field pinelands provide habitat to gopher tortoises (*Gopherus polyphemus*), Bachman's sparrows (*Aimophila aestivalis*), fox squirrels (*Sciurus niger*), and reintroduced red-cockaded woodpeckers (*Picoides borealis*) (Masters et al. 2003). Even so, old-field successional habitats are thought to constitute a major shift in floristic characteristics from their original composition (Means and Grow 1985, Myers 1990). In the eastern portion of the southeastern U.S. Coastal Plain, upland old-field

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pinelands once dominated by longleaf pine (*Pinus palustris*) and wiregrass (*Aristida stricta*) are now typically dominated by loblolly pine (*Pinus taeda*) and shortleaf pine (*P. echinata*) (Moser et al. 2002) and have lost wiregrass as a component of their herbaceous ground cover (Crafton and Wells 1934, McQuilken 1940, Oosting 1942, Grelen 1962, Clewell 1986, Hedman et al. 2000). It is generally observed that old-field vegetation has a smaller number of native species and that these species are relatively common (Clewell 1986, Hedman et al. 2000, Kirkman et al. 2004). This prevalence is attributed to their ability to disperse to and colonize disturbed soil (Kirkman et al. 2004). Thus, old-field pinelands may be placed at a lower priority for conservation and protection.

Although species composition of native longleaf pine habitats has often been described (e.g., Bridges and Orzell 1989, Hardin and White 1989, Drew et al. 1998, Kush and Meldahl 2000, Varner et al. 2003, Carter et al. 2004), comparatively few published studies have compared vegetation composition between old-field and relatively undisturbed upland pine habitats (Hedman et al. 2000, Kirkman et al. 2004). These studies have shown significant differences in species composition between old-field sites and "reference sites" representing the native community type, as well as lower species richness in old-field sites because of dispersal limitations on certain native species (Kirkman et al. 2004). Native pineland communities are also sensitive to vehicular traffic, soil compaction, and surface scarification associated with planted pine site preparation, logging, and military training (Hedman et al. 2000, Dale et al. 2002, Smith et al. 2002). However, soil disturbance associated with intensive agriculture appears to have the greatest impact on species composition in fire-maintained pinelands relative to these other disturbance types (Hedman et al. 2000).

The sensitivity of many species to anthropogenic disturbance makes them useful indicators of habitat integrity (Noss 1990, Kimberling et al. 2001, McLachlan and Bazely 2001, Moffatt and McLachlan 2004). Indicator species should be sensitive to the environmental stress of interest, and thus may indicate the biological integrity of an ecosystem (Dale and Beyeler 2001) and abundant and tractable components of the system (Welsh and Ollivier 1998). Indicator species for relatively undisturbed native pinelands maintained with frequent fire in the Coastal Plain have been proposed in other studies (Rodgers and Provencher 1999, Dale et al. 2002, Smith et al. 2002, Kirkman et al. 2004). However, additional studies are needed both to confirm the reliability of certain widely distributed species as indicators throughout the region as well as to identify local indicator species among physiographic features within the region (e.g., Dougherty Plain, Tifton Uplands).

The goals of this study were to characterize and contrast the species composition between native and old-field upland clayhill pinelands in Southwest Georgia, to test the utility of these data in interpreting the community integrity of sites with unknown disturbance history, and to identify possible indicator species that can aid in distinguishing these community types. The results should

assist in identifying natural areas for conservation and appropriate management and for assessing the ecological value and restoration potential for old-field pine forests managed with frequent fire.

METHODS

Study Area

The study was conducted within the Red Hills region of southern Georgia and northern Florida on Pebble Hill Plantation (PHP) (30°35'N, 84°20'W), which covered approximately 1,222 ha in Grady and Thomas counties, Georgia. The Red Hills region was characterized by gently sloping, well-drained sandy or loamy soils underlain by clayey or sandy sub-horizons (Calhoun 1979). Plots were more or less evenly distributed among the following soil types: Bonneau loamy sand, Dothan loamy sand, Lucy loamy sand, Tifton loamy sand, Faceville sandy loam, and Nankin-Cowarts sandy loam. Mean annual temperature was 19.6°C (11°–27.4°C monthly means) and mean annual precipitation was 1,373.4 mm (Southeast Regional Climate Center 2004). The growing season for this region was from early March to November (Calhoun 1979; T.E. Ostertag and K.M. Robertson, unpublished data). PHP had been managed for northern bobwhite (*Colinus virginianus*) and timber during the past century, primarily with the use of frequent prescribed fire (1- to 2-y fire interval) and both even- and uneven-aged management systems.

The forested upland habitats at PHP were a mixture of old fields, pine plantations, and native pinelands (never plowed) (Robertson and Ostertag 2003). The native areas had a canopy dominated by longleaf pine, often mixed with shortleaf pine, and an understory supporting a high diversity of other woody plants, forbs, and grasses, especially wiregrass. The old-field pineland habitats had a canopy dominated by some mixture of shortleaf and loblolly pines and an understory of woody species typical of disturbed areas in the region, such as water oak (*Quercus nigra*) and sweetgum (*Liquidambar styraciflua*), and herbs composed of some subset of native species as well as agricultural weeds. Timber management had been similarly applied to the specifically selected native and old-field pineland sites, such that stand densities and structure were similar between the two habitat types (average of 9.6 ± 6.3 SD m²/ha in native stands based on 54 random plots, $10.9 \text{ m}^2 \pm 6.3 \text{ m}^2$ basal area/ha in old-field stands based on 59 random plots [K.M. Robertson, unpublished data]).

Site Selection and Sampling Methods

Plant communities were compared among the following cover types: native ground cover, old fields, and areas of unknown soil disturbance history. Native ground cover was identified by the presence of wiregrass, based on the general observation that wiregrass does not readily return to extensive areas of heavily disturbed soil or prolonged fire suppression (Hebb

1957, 1971; Woods 1959; Grelen 1962; Harris et al. 1974; Schultz and Wilhite 1974; White et al. 1975; Myers 1990). Old-field sites were identified as cultivated in a 1928 (earliest known) aerial photograph. Based on the age of current pines and other historical records, we estimated these sites had been abandoned from cultivation circa 1950. Areas of unknown management history (may or may not be native ground cover) were forest at the time of the 1928 aerial photograph (current age of trees approximately 150 y old) but lacked wiregrass at the time of this study.

For each of the three cover types, 12 permanent sampling plots were established, with 3 plots in each of 4 separate burn units. Burn units ranged in area from 2.6 to 18.5 ha. Burn units and plot locations within burn units were randomly chosen using ArcView 3.2 Animal Movement extension (Environmental Systems Research Institute 1999). Potential plot locations were limited to ridge tops and the upper halves of slopes to restrict them to upland pine habitats. Each plot was 100 m² (10 m × 10 m), with nested subplots of 10 m² and 1 m² in one corner. Thus, the 36 total plots covered 3,600 m². Species were censused starting with the 1-m² subplot, then proceeding to the 10-m² and the 100-m² plots. Species of all herbaceous vascular plants and woody shrubs and trees <1 m high were censused within the plots. Percent cover within the 100-m² plot was estimated for each species using a modified Daubenmire cover class method (Peet et al. 1998). Maximum cover for each species was 100%, but cumulative cover for all species could exceed 100%. The plots were censused four times between fall 2003 and spring 2005 (once in each of the following: October–November 2003, October–November 2004, April–May 2004, and April–May 2005) to incorporate seasonal and annual variation in the presence and visibility of plant species. Plants were identified to species and otherwise were not included in the analysis. Unidentified plants were generally seedlings or those badly damaged by herbivory or senescence. Unidentified specimens accounted for 8% of all samples collected.

Species composition in the study plots was compared to species lists compiled between July 1995 and November 2002 from transects running throughout PHP (A. Gholson and C. Martin, PHP, unpublished data). Transects used in the analysis were limited to those entirely contained within native ($n = 9$) and old-field ($n = 8$) cover types. There were no transects contained entirely within the unknown-disturbance-history cover type. These censuses incorporated plants that could be easily observed by the surveyors to either side of each transect (total width of 4 m). Transects varied in length from 160 to 719 m due to size of the burn unit. The total area censused for all transects was approximately 22,000 m². Transects were not bound by topographic or hydrological constraints as were the study plots, such that lower hillslope and wetland pine-lands were included. All plants, including trees and shrubs regardless of size, were recorded. Taxonomy and nomenclature for both the plot censuses and transects follows Wunderlin and Hansen (2003).

Data Analysis

A detrended canonical correspondence analysis (DCCA) in CANOCO 4.5 (ter Braak and Smilauer 2002) was used to test the null hypothesis of no predictable difference in species composition and cover among native, old-field, or unknown study plots. The median cover values for 100-m² plots estimated for species were used as the response variables in the analysis, and plots served as units of replication. Monte Carlo permutations (499 iterations) on the first canonical axis were used to produce an F -statistic testing for differences among sampling plots (Leps and Smilauer 2003). A detrended correspondence analysis (DCA) in CANOCO 4.5 was used to create an ordination diagram in CANODRAW 4.5 (ter Braak and Smilauer 2002) to visually assess floristic similarity among sampling plots (Leps and Smilauer 2003). Similarly, DCCA and DCA were used to test for differences in species composition between native and old-field transects based on presence–absence of species.

Binomial analyses were used to identify the most likely indicator species for native and old-field ground cover using both the plot and transect data. Plots (excluding those of unknown disturbance history) containing a given species were assigned a binomial variable based on whether they were native or old field. Binomial tables were used to determine if there was a statistically nonrandom (two-sided $\alpha = 0.05$) association of a species with a particular ground-cover type (Sokal and Rohlf 1995). Each occurrence of a species in a plot or transect was considered to have a 0.5 probability of occurring in either native or old-field cover by chance; for example, occurrence of a species in 5 native plots and none in old-field plots would be assigned the probability of 0.031 (nonrandom). Indicators were defined as plants characteristic of and confined predominately to a particular habitat based on the binomial analysis. This assessment was made separately for plots and transects.

Plant species affiliated with one cover type were placed in one of three categories: 1) restricted to one cover type and in a sufficient number of plots or transects (5 or more) to have a statistically significant affinity ($P \leq 0.031$); 2) found predominately in one cover type with a statistically significant affinity; 3) found only in one cover type in 3 or 4 plots or transects ($P = 0.125$ or 0.063, respectively). Category 3 was designed to identify plants with a trend toward a definitive association with one ground-cover type versus the other, which might be revealed with additional sampling. Species for which plot and transect data contradicted one another were not listed as potential indicators.

RESULTS

A total of 232 plant species in 53 families were identified in the study plots, of which six were non-native species (Appendix A). Five families, Poaceae, Fabaceae, Asteraceae, Fagaceae, and Ericaceae, encompassed 58% of the species, and 31 families were represented by a single species. The number of species

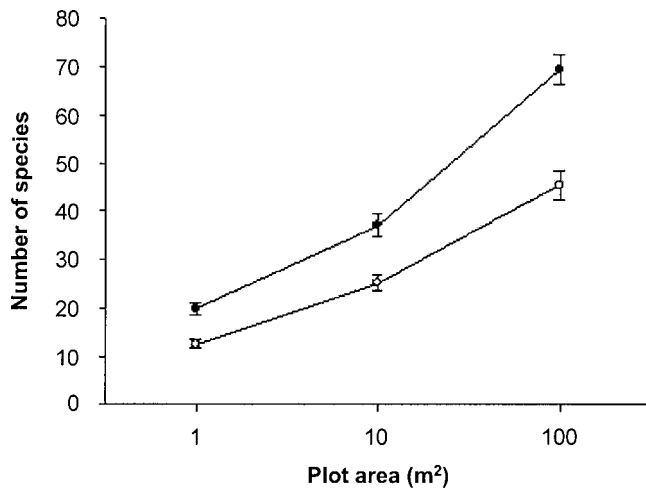


Fig. 1. Species area curve for vegetation plot data collected spring 2003–fall 2005 on Pebble Hill Plantation, Grady and Thomas counties, Georgia. Mean number of species and standard error among 12 plots are given for each scale censused within 100-m² plots. Closed circles indicate native ground cover and open circles indicate old fields.

restricted to one cover type was 56 for native, 15 for old-field, and 16 for unknown-disturbance-history cover plots. Native and old-field plots shared in common 103 species (Appendix A). Average species richness among the 100-m² plots was 69.5 ± 15.9 (\pm SD) for native plots, 45.4 ± 10.8 in old-field plots, and 46.4 ± 10.8 in unknown-disturbance-history plots. Average species richness was approximately 50% higher in native compared to old-field cover at each of the three spatial scales surveyed (Figure 1).

The transect data yielded a total of 432 plant species in 92 families, of which 36 were nonnative species (Appendix A). Five families, Asteraceae, Fabaceae, Poaceae, Cyperaceae, and Fagaceae, contained 49% of the species, and 42 families were represented by a single species. There were 111 species confined to native ground-cover transects and 111 species confined to old-field transects, and 210 species were in both native and old-field transects. Of species confined to native transects, 69% occurred in only 1 of 9 transects, 15% occurred in 2 transects, and 16% occurred in >2 . Of species confined to old fields, 66% occurred in 1 of 8 transects, 29% occurred in 2, and 5% occurred in >2 .

Native and old-field plots were distinctly different in their species composition, as indicated by the Monte Carlo test ($F = 3.827$, $P = 0.002$) and DCA scatter plot (Figure 2). Unknown-disturbance-history plots were largely clustered with old-field plots, although a few appeared to be in a transition zone between old-field and native plots (Figure 2). Similarly, species composition differed significantly between native ground-cover and old-field transects ($F = 1.311$, $P = 0.008$) and appeared to be distinctly different in the scatter plot (Figure 3).

Native sites were characterized by greater dominance of grasses, especially wiregrass and slender bluestem (*Schizachyrium tenerum*), and a lower dominance of forbs and woody species compared to old-field sites (Table 1). Native plots had an average ab-

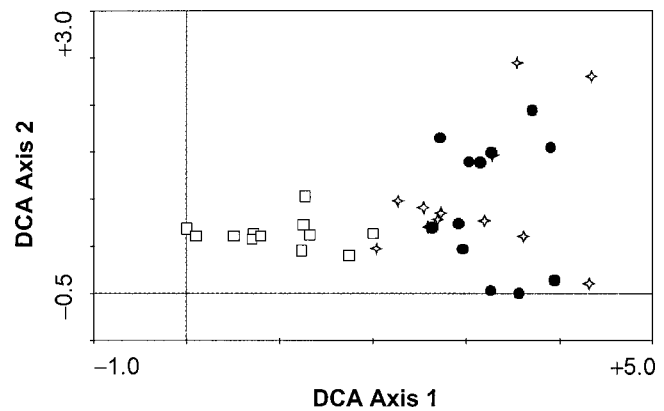


Fig. 2. Ordination diagram of vegetation plot data using detrended correspondence analysis (DCA). Data were collected between spring 2003 and fall 2005 on Pebble Hill Plantation, Grady and Thomas counties, Georgia. Squares indicate native ground-cover plots, circles indicate old-field plots, and stars indicate plots of unknown soil disturbance history. Proximity of symbols reflects their similarity in species composition and cover.

solute cover of 29.6% grasses, 7.0% forbs, and 38.4% woody shrubs and vines. In comparison, old-field sites had an absolute cover of 0.6% grasses, 15.7% forbs, and 63.2% woody shrubs and vines (Table 1). Of the forbs in old-field plots, several species (hyssopleaf thoroughwort [*Eupatorium hyssopifolium*], dogfennel [*E. capillifolium*], lesser snakeroot [*Ageratina aromatica*], and wrinkleleaf goldenrod [*Solidago rugosa*]) are structurally similar to woody shrubs and account for approximately two-thirds of forb cover (Table 1).

Potential indicator species were identified for each cover type by using binomial analyses in both the plot and transect data (Table 2). Twelve species were restrict-

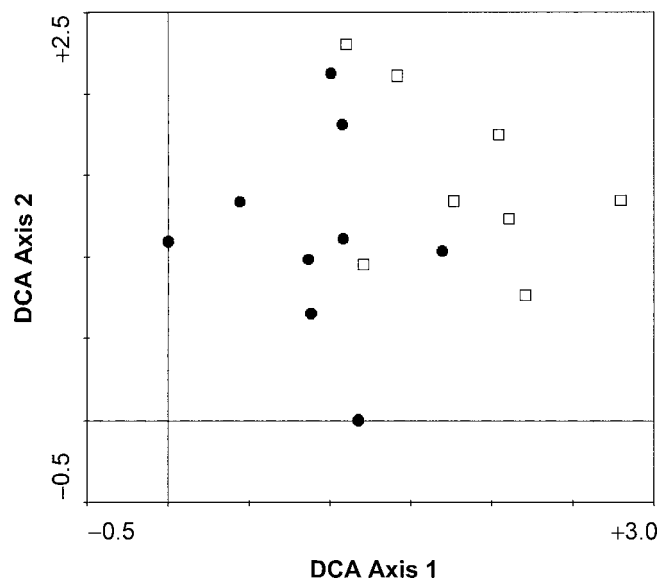


Fig. 3. Ordination diagram of vegetation transect data using detrended correspondence analysis (DCA). Data were collected between spring 2003 and fall 2005 on Pebble Hill Plantation, Grady and Thomas counties, Georgia. Squares indicate native ground-cover transects and circles indicate old-field transects. Proximity of symbols reflects their similarity based on species presence-absence.

Table 1. The 30 most dominant plant species, based on percent absolute cover, for native and old-field plots, in descending order of average cover among study plots, Pebble Hill Plantation, Grady and Thomas counties, Georgia, from spring 2003 to fall 2005.

Native ground-cover sites			Old-field sites		
Species	Growth form ^a	% cover	Species	Growth form ^a	% cover
<i>Aristida stricta</i>	G	24.3	<i>Liquidambar styraciflua</i>	W	14.8
<i>Ilex glabra</i>	W	7.3	<i>Callicarpa americana</i>	W	10.5
<i>Pinus palustris</i>	W	6.2	<i>Pinus taeda</i>	W	7.8
<i>Schizachyrium tenerum</i>	G	4.0	<i>Eupatorium hyssopifolium</i>	F	4.6
<i>Quercus elliotii</i>	W	3.8	<i>Rubus cuneifolius</i>	W	3.9
<i>Quercus incana</i>	W	3.8	<i>Pinus echinata</i>	W	3.6
<i>Quercus laurifolia</i>	W	2.9	<i>Pityopsis aspera</i>	F	3.5
<i>Quercus marilandica</i>	W	2.4	<i>Myrica cerifera</i>	W	2.8
<i>Quercus falcata</i>	W	2.0	<i>Rhus copallinum</i>	W	2.7
<i>Pteridium aquilinum</i>	F	1.9	<i>Quercus falcata</i>	W	2.6
<i>Vaccinium corymbosum</i>	W	1.7	<i>Quercus nigra</i>	W	2.4
<i>Quercus stellata</i>	W	1.2	<i>Prunus serotina</i>	W	2.1
<i>Viola palmata</i>	F	1.1	<i>Ageratina aromatica</i>	F	2.0
<i>Vaccinium darrowii</i>	W	1.0	<i>Eupatorium capillifolium</i>	F	1.8
<i>Rhus copallinum</i>	W	0.9	<i>Quercus stellata</i>	W	1.5
<i>Myrica cerifera</i>	W	0.8	<i>Vitis rotundifolia</i>	V	1.5
<i>Pinus echinata</i>	W	0.8	<i>Erythrina herbacea</i>	F	1.3
<i>Pityopsis aspera</i>	F	0.8	<i>Carya alba</i>	W	1.2
<i>Vaccinium myrsinites</i>	W	0.7	<i>Smilax glauca</i>	V	1.1
<i>Mimosa quadrivalvis</i>	F	0.7	<i>Quercus virginiana</i>	W	1.0
<i>Rhynchosia reniformis</i>	F	0.6	<i>Diospyros virginiana</i>	W	1.0
<i>Sorghastrum nutans</i>	G	0.6	<i>Lespedeza virginica</i>	F	0.9
<i>Dyschoriste oblongifolia</i>	F	0.6	<i>Quercus laurifolia</i>	W	0.7
<i>Gelsemium sempervirens</i>	V	0.6	<i>Rubus argutus</i>	W	0.6
<i>Rubus cuneifolius</i>	W	0.6	<i>Chamaecrista nictitans</i>	F	0.6
<i>Seymeria pectinata</i>	F	0.5	<i>Saccharum alopecuroides</i>	G	0.6
<i>Symphytotrichum concolor</i>	F	0.4	<i>Solidago rugosa</i>	F	0.5
<i>Carya alba</i>	W	0.4	<i>Hypericum hypericoides</i>	F	0.5
<i>Cornus florida</i>	W	0.4	<i>Ampelopsis arborea</i>	V	0.4
<i>Vaccinium arboreum</i>	W	0.4	<i>Parthenocissus quinquefolia</i>	V	0.4

^a Growth form: F, forb; G, grass; V, woody vine; W, woody shrub or tree.

ed to native ground-cover plots with a sufficiently high frequency to be statistically associated with this cover type (category 1), 4 species were statistically associated but were not restricted to native plots (category 2), and 11 species were found only in native cover in 3 or 4 plots (category 3) (Table 2). Old fields had 3 species in category 1, 2 species in category 2, and 0 species in category 3 (Table 2). Native ground-cover transects had 5 indicator species in category 1, 0 species in category 2, and 12 species in category 3. Old-field transects had 0 species in categories 1 or 2 and 8 species in category 3 (Table 2). Considering plots and transects together, a total of 17 species were identified as having statistically significant associations with native ground cover and 5 species had significant associations with old fields (Table 2). An additional 16 species occurring only in native ground cover and 8 species occurring only in old fields had affinities to those cover types, with random probabilities of <0.125.

DISCUSSION

Results from both the study plots and transect data showed distinctive compositional differences between native and old-field sites. In the study plots, this difference was mostly attributable to the larger number of species occurring on native sites and differences in relative cover of species. Old fields were primarily composed of a subset of native species, with only a

few exceptions. Additionally, nonnative species were found in greater numbers in old-field plots and transects than in native ground-cover plots and transects. These results are consistent with other studies comparing old-field and native sites (Means and Grow 1985, Myers 1990, Hedman et al. 2000, Kirkman et al. 2004). Limitations on dispersal and ability for native species to colonize disturbed sites are likely the strongest determinants of species differences between the community types (Kirkman et al. 2004). It should be noted that in our and in the previously cited studies, native pinelands were located on the same properties as the old fields studied, apparently providing a source for colonizing propagules. The proximity of native sites may be a key factor in direction of succession of abandoned old fields (Kirkman et al. 2004).

The plots of unknown soil disturbance history were generally similar to old-field plots in their species composition and dominance, with the exception of some plots that appeared to be in a transition zone between the two. The latter plots may have been subject to various degrees of disturbances other than intensive agriculture, such as those associated with logging or mechanical methods of shrub control. Given the history of selective timber management on our particular study sites, these areas may have been affected to some degree by past vehicular traffic and surface disturbance. The other plots we interpret as being old fields that were abandoned before the earliest available

Table 2. Potential indicator species of native ground cover and old fields from Pebble Hill Plantation, Grady and Thomas counties, Georgia, 2003–2005, and their identification as suggested indicator species in other studies in the southeastern U.S. Coastal Plain. Nomenclature follows Wunderlin and Hansen (2003).

Species	Category ^a		
	Plots	Transects	Other studies ^b
Native ground-cover indicators			
<i>Angelica dentata</i>	3	1	F
<i>Aristida stricta</i>	— ^c	1	B, D, E
<i>Asimina angustifolia</i>	2		
<i>Aureolaria pedicularia</i>	3		
<i>Carphephorus odoratissimus</i>	1	3	
<i>Chrysopsis mariana</i>		1	D, E
<i>Dalea carnea</i>	3		
<i>Dyschoriste oblongifolia</i>	1		B, E
<i>Euphorbia curtisii</i>		3	A
<i>Euphorbia discoidalis</i>	1		A
<i>Gaylussacia dumosa</i>	1	3	
<i>Gaylussacia frondosa</i>	3	1	
<i>Gymnopogon brevifolius</i>	3		
<i>Helianthus angustifolius</i>	3		
<i>Hypericum hypericoides</i>		3	
<i>Ilex glabra</i>	1		
<i>Lobelia amoena</i>		3	
<i>Mimosa quadrivalvis</i>	2		E
<i>Monotropa uniflora</i>	3		
<i>Muhlenbergia capillaris</i>	3	3	E
<i>Panicum virgatum</i>		3	E
<i>Pityopsis graminifolia</i>	1		
<i>Pleopeltis polypodioides</i>		3	
<i>Pteridium aquilinum</i>	1		B, C, E
<i>Quercus elliotii</i>	1		
<i>Quercus incana</i>	1	3	
<i>Quercus laevis</i>	1	3	
<i>Sebastiania fruticosa</i>		3	
<i>Saccharum coarctatum</i>	3		
<i>Salvia azurea</i>	1	1	
<i>Seymeria pectinata</i>		3	
<i>Strophostyles umbellata</i>	2		C, E, F
<i>Stylisma patens</i>	3		
<i>Symphotrichum adnatum</i>	1		B, E
<i>Symphotrichum concolor</i>	2		C, E
<i>Tephrosia virginiana</i>	3		A, B, C, E, F
Old-field indicators			
<i>Croton glandulosus</i>		3	
<i>Erythrina herbacea</i>	2		
<i>Eupatorium capillifolium</i>	2		
<i>Gamochaeta pensylvanica</i>		3	
<i>Lespedeza procumbens</i>		3	
<i>Liatris tenuifolia</i>		3	
<i>Quercus virginiana</i>	1		
<i>Ruellia caroliniensis</i>		3	A, B, D
<i>Sebastiania fruticosa</i>		3	
<i>Seymeria pectinata</i>		3	
<i>Solidago rugosa</i>	1		
<i>Trichostema dichotomum</i>	1		
<i>Yucca filamentosa</i>		3	

^a Category: 1, plants restricted to one cover type with a statistical affinity; 2, plants predominately in one cover type with a statistical affinity; 3, plants restricted to one cover type without statistical affinity.

^b Other studies: A, Rodgers and Provencher 1999; B, Hedman et al. 2000; C, Dale et al. 2002; D, Smith et al. 2002; E, Kirkman et al. 2004; F, Carter et al. 2004.

^c *Aristida stricta* was used as an indicator of whether or not sites had been disturbed and hence was not included in the analysis.

aerial photograph, based on their close similarity in species composition and structure to confirmed old-field plots. This assessment provides an example of the way in which baseline data characterizing the composition and dominance of native and old-field sites may be used in interpreting the disturbance history of pineland ground cover through multivariate analysis.

In contrast to the study plots, transects had an equal number of species limited to either native or old-field sites. This difference is attributable to the incorporation of a much larger area and thus a greater accumulation of rare species in transects relative to the plots, reflected by the fact that the great majority of species in the transect data occurred in only one or two transects. These species may be either naturally rare species or species associated with isolated disturbances; in either case, they are likely to be missed by transects in the other cover type.

In addition to compositional differences between native and old-field sites, there were significant differences in structure and species dominance. In particular, native sites were characterized by high grass cover and relatively lower forb and woody species cover, whereas old fields had only trace cover by grass and were dominated by large-statured forbs and woody plants. Clearly the likelihood of dispersal and establishment of plants from native areas to abandoned fields is dependent, at least in part, on whether they are grass, forb, or woody species. The association of higher forb and woody species dominance with degree of soil disturbance in southern pinelands has been noted in other studies (Hedman et al. 2000, Dale et al. 2002, Smith et al. 2002). This observation has implications for the management techniques that may be required to restore certain old-field pinelands to a more natural vegetation structure, including more frequent burning (White et al. 1991, Waldrop et al. 1992, Provencher et al. 2001c, Glitzenstein et al. 2003) and use of selective herbicides (Brockway et al. 1998, Litt et al. 2001, Provencher et al. 2001a,b, Jones and Chamberlain 2004, Miller and Miller 2004).

In addition to distinguishing native versus old-field pinelands, our study reveals the potential for old-field pine forests to harbor a significant proportion of native pineland species when managed with frequent fire and selective timber thinning. Such forests are more appropriately described as partially to mostly restored native pinelands rather than purely anthropogenic habitats, as demonstrated by their potential to support a diversity of rare animal species adapted to native pinelands (Masters et al. 2003, Provencher et al. 2003). These qualities attest to the benefits of frequent fire and appropriate timber management for restoring and maintaining high-quality wildlife habitat in post-agriculture pine forests.

We propose 17 species as indicators of native ground cover lacking intensive soil disturbance, based on their statistical affinity to that cover type, and 16 species to be potential indicators based on their nearly significant associations with native sites. Of these, 14 species have been identified as indicators of native ground cover lacking an intensive soil disturbance his-

tory in Coastal Plain pinelands in other studies (Rodgers and Provencher 1999, Hedman et al. 2000, Dale et al. 2002, Smith et al. 2002, Kirkman et al. 2004; Table 2). In addition to wiregrass, used to identify native plots in this study and thus excluded from analysis, coastalplain angelica (*Angelica dentata*), Maryland goldenaster (*Chrysopsis mariana*), oblongleaf snakeherb (*Dyschoriste oblongifolia*), Curtis' spurge (*Euphorbia curtisii*), summer spurge (*Euphorbia discoidalis*), sensitive brier (*Mimosa quadrivalvis*), hairawn muhly (*Muhlenbergia capillaris*), switchgrass (*Panicum virgatum*), brackenfern (*Pteridium aquilinum*), pink fuzzybean (*Strophostyles umbellata*), eastern silver aster (*Symphyotrichum concolor*), and goat's rue (*Tephrosia virginiana*) appear to have geographically broad utility as indicators of pineland sites which have experienced minimal ground-cover disturbance. Other species, live oak (*Quercus virginiana*), Carolina wild petunia (*Ruellia caroliniensis*), wrinkleleaf goldenrod, and forked bluecurls (*Trichostema dichotomum*), appear to be reliable indicators of soil disturbance.

These studies of upland pine forests also point out the variation in floristic composition among pinelands. In individual studies, most of the species identified as potential indicators have not been found in sufficiently large numbers in other studies to merit distinction (Hedman et al. 2000, Dale et al. 2002, Smith et al. 2002, Kirkman et al. 2004). This may be related to inherent site and soil differences (i.e., sandy vs. clayey soils). However, given that it is recommended to use the full suite of plant species available in interpreting natural habitat integrity (Zonneveld 1983, Kremen 1992, Carignan and Villard 2002), all suggested species, when appearing in significant numbers, should be considered as indicative of native community presence. Such interpretations are critical and time sensitive, given the ongoing conversion of native pinelands to other land uses in the southeastern U.S. Coastal Plain.

In summary, southern pineland native ground cover contains a significant number of species that do not readily become reestablished during several years or even decades following extensive soil disturbance. The extreme sensitivity of many native species to intensive soil disturbance underscores the need to identify remaining undisturbed areas for conservation, as facilitated by the use of indicator species suggested by this and other studies. Nevertheless, our data demonstrate the potentially high species richness of native plants in old fields managed with frequent fire and selective timber management. Understanding the differences and similarities between these community types, including species composition and community structure, is important for prioritizing future conservation, restoration, and management of the species they support.

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Appendix A. Plant species list for plots and transects, with number of occurrences for three ground-cover types, Pebble Hill Plantation, Grady and Thomas counties, Georgia, 2003–2005. Nomenclature follows Wunderlin and Hansen (2003). Asterisks denote nonnative species.

Species	Family	Plots ^a			Transects ^a	
		N	O	U	N	O
<i>Acalypha gracilens</i>	Euphorbiaceae	3	2	0	4	5
<i>Acer rubrum</i>	Sapindaceae	2	1	3	3	4
<i>Aesculus pavia</i>	Sapindaceae	0	0	0	0	1
<i>Agalinis divaricata</i>	Orobanchaceae	2	0	0	0	0
<i>Agalinis fasciculata</i>	Orobanchaceae	0	0	0	1	1
<i>Agalinis filifolia</i>	Orobanchaceae	1	0	0	0	0
<i>Agalinis tenuifolia</i>	Orobanchaceae	0	0	0	1	0
<i>Ageratina aromatica</i>	Asteraceae	5	3	8	5	8
<i>Ageratina jucunda</i>	Asteraceae	0	0	0	1	3
<i>Agrimonia microcarpa</i>	Rosaceae	0	0	0	1	0
<i>Albizia julibrissin*</i>	Fabaceae	0	0	0	0	1
<i>Aletris aurea</i>	Nartheciaceae	0	0	0	1	0
<i>Aletris obovata</i>	Nartheciaceae	1	0	0	0	0
<i>Ambrosia artemisiifolia</i>	Asteraceae	0	4	0	4	7
<i>Ampelopsis arborea</i>	Vitaceae	0	3	2	0	1
<i>Amphicarpaea bracteata</i>	Fabaceae	1	0	0	0	0
<i>Andropogon gerardii</i>	Poaceae	2	0	0	0	0
<i>Andropogon gyrans</i>	Poaceae	1	1	0	2	1
<i>Andropogon longiberbis</i>	Poaceae	2	0	0	0	0
<i>Andropogon ternarius</i>	Poaceae	3	0	1	1	1
<i>Andropogon virginicus</i>	Poaceae	2	4	5	2	3
<i>Angelica dentata</i>	Apiaceae	4	0	0	5	0
<i>Anthaenaria villosa</i>	Poaceae	0	0	0	1	0
<i>Apios americana</i>	Fabaceae	0	0	0	0	1
<i>Apocynum cannabinum</i>	Apocynaceae	0	0	0	0	1
<i>Aralia spinosa</i>	Araliaceae	0	0	0	1	3
<i>Arisaema triphyllum</i>	Araceae	0	0	0	0	2
<i>Aristida purpurascens</i>	Poaceae	3	4	2	1	0
<i>Aristida stricta</i>	Poaceae	12	0	3	7	3
<i>Aristolochia serpentaria</i>	Aristolochiaceae	2	0	0	0	0
<i>Arundinaria gigantea</i>	Poaceae	0	0	0	0	2
<i>Asclepias amplexicaulis</i>	Apocynaceae	0	0	0	0	1
<i>Asclepias cinerea</i>	Apocynaceae	1	0	0	0	0
<i>Asclepias tuberosa</i>	Apocynaceae	1	0	1	0	0
<i>Asclepias variegata</i>	Apocynaceae	0	0	0	0	1
<i>Asclepias verticillata</i>	Apocynaceae	0	0	0	2	0
<i>Asimina angustifolia</i>	Annonaceae	8	1	3	6	7
<i>Asimina parviflora</i>	Annonaceae	0	0	0	1	0
<i>Asplenium platyneuron</i>	Aspleniaceae	0	0	0	1	4
<i>Athyrium filix-femina</i>	Dryopteridaceae	0	0	0	0	2
<i>Aureolaria flava</i>	Orobanchaceae	0	0	0	1	2
<i>Aureolaria pedicularia</i>	Orobanchaceae	4	0	0	2	0
<i>Aureolaria virginica</i>	Orobanchaceae	0	0	0	1	0
<i>Baccharis halimifolia</i>	Asteraceae	0	0	0	2	1
<i>Baptisia lecontei</i>	Fabaceae	0	0	0	2	0
<i>Bidens bipinnata</i>	Asteraceae	0	0	0	0	1
<i>Bignonia capreolata</i>	Bignoniaceae	0	0	1	1	4
<i>Boehmeria cylindrica</i>	Urticaceae	0	0	0	0	1
<i>Brickellia eupatorioides</i>	Asteraceae	0	0	0	1	0
<i>Buchnera americana</i>	Orobanchaceae	0	0	0	1	1
<i>Callicarpa americana</i>	Lamiaceae	7	10	12	6	9
<i>Campsis radicans</i>	Bignoniaceae	0	0	0	2	7
<i>Carex abscondita</i>	Cyperaceae	0	0	0	0	1
<i>Carex comosa</i>	Cyperaceae	0	0	0	0	1
<i>Carex retroflexa</i>	Cyperaceae	0	0	0	0	1
<i>Carex verrucosa</i>	Cyperaceae	0	0	0	1	0
<i>Carphephorus corymbosus</i>	Asteraceae	0	0	0	1	0
<i>Carphephorus odoratissimus</i>	Asteraceae	6	0	0	4	0
<i>Carpinus caroliniana</i>	Betulaceae	0	0	0	0	1
<i>Carya alba</i>	Juglandaceae	5	5	7	8	6
<i>Carya glabra</i>	Juglandaceae	3	1	1	2	3
<i>Carya illinoensis*</i>	Juglandaceae	0	0	0	0	2
<i>Castanea pumila</i>	Fagaceae	0	0	1	0	1
<i>Ceanothus americanus</i>	Rhamnaceae	2	0	0	2	3

Appendix A. Continued.

Species	Family	Plots ^a			Transects ^a	
		N	O	U	N	O
<i>Ceanothus microphyllus</i>	Rhamnaceae	0	0	0	2	0
<i>Centrosema virginianum</i>	Fabaceae	4	2	1	4	3
<i>Cercis canadensis</i>	Fabaceae	0	0	0	0	2
<i>Chamaecrista fasciculata</i>	Fabaceae	0	0	0	3	3
<i>Chamaecrista nictitans</i>	Fabaceae	9	10	7	4	2
<i>Chamaesyce hirta</i>	Euphorbiaceae	0	0	0	0	1
<i>Chasmanthium laxum</i>	Poaceae	0	2	4	4	5
<i>Chionanthus virginicus</i>	Oleaceae	0	0	0	1	1
<i>Chrysopsis mariana</i>	Asteraceae	7	2	2	7	0
<i>Cirsium horridulum</i>	Asteraceae	0	0	0	1	0
<i>Clethra alnifolia</i>	Clethraceae	0	0	0	1	0
<i>Clitoria mariana</i>	Fabaceae	6	2	4	3	3
<i>Cnidioscolus stimulosus</i>	Euphorbiaceae	5	3	1	3	6
<i>Commelina erecta</i>	Commelinaceae	0	0	0	1	2
<i>Conoclinium coelestinum</i>	Asteraceae	3	4	0	4	4
<i>Conyza bonariensis</i>	Asteraceae	0	0	0	0	1
<i>Conyza canadensis</i>	Asteraceae	1	1	0	2	2
<i>Cornus florida</i>	Cornaceae	5	2	8	5	7
<i>Crataegus flava</i>	Rosaceae	0	0	0	2	1
<i>Crataegus uniflora</i>	Rosaceae	0	0	1	0	0
<i>Croptilon divaricatum</i>	Asteraceae	0	0	0	1	0
<i>Crotalaria purshii</i>	Fabaceae	0	0	0	2	1
<i>Crotalaria rotundifolia</i>	Fabaceae	3	2	1	2	3
<i>Crotalaria spectabilis*</i>	Fabaceae	0	2	0	0	2
<i>Croton argyranthemus</i>	Euphorbiaceae	0	0	0	0	1
<i>Croton glandulosus</i>	Euphorbiaceae	0	0	0	0	4
<i>Cuphea carthagenensis*</i>	Lythraceae	0	0	0	1	2
<i>CyclospERMUM leptophyllum</i>	Apiaceae	0	0	0	0	2
<i>Cynodon dactylon*</i>	Poaceae	0	0	0	1	0
<i>Cyperus croceus</i>	Cyperaceae	0	0	0	1	2
<i>Cyperus filiculmis</i>	Cyperaceae	1	0	0	0	0
<i>Cyperus hystricinus</i>	Cyperaceae	0	0	0	1	0
<i>Cyperus plukenetii</i>	Cyperaceae	1	2	0	2	1
<i>Cyperus retrorsus</i>	Cyperaceae	0	0	0	1	2
<i>Cyrilla racemiflora</i>	Cyrtillaceae	0	0	0	1	1
<i>Dalea carnea</i>	Fabaceae	4	0	1	2	0
<i>Dalea pinnata</i>	Fabaceae	0	0	0	2	2
<i>Decumaria barbara</i>	Hydrangeaceae	0	0	0	0	1
<i>Desmodium ciliare</i>	Fabaceae	5	4	3	1	1
<i>Desmodium floridanum</i>	Fabaceae	1	0	1	1	4
<i>Desmodium glabellum</i>	Fabaceae	1	0	0	0	0
<i>Desmodium laevigatum</i>	Fabaceae	0	2	3	3	5
<i>Desmodium lineatum</i>	Fabaceae	1	1	1	4	1
<i>Desmodium marilandicum</i>	Fabaceae	1	0	0	4	3
<i>Desmodium nudiflorum</i>	Fabaceae	0	0	0	1	1
<i>Desmodium obtusum</i>	Fabaceae	0	0	0	1	0
<i>Desmodium paniculatum</i>	Fabaceae	0	3	1	3	5
<i>Desmodium perplexum</i>	Fabaceae	1	0	0	0	0
<i>Desmodium rotundifolium</i>	Fabaceae	0	0	0	2	0
<i>Desmodium strictum</i>	Fabaceae	1	0	1	1	0
<i>Desmodium tenuifolium</i>	Fabaceae	1	0	0	2	0
<i>Desmodium viridiflorum</i>	Fabaceae	1	0	1	4	4
<i>Dichanthelium aciculare</i>	Poaceae	4	3	1	3	1
<i>Dichanthelium acuminatum</i>	Poaceae	1	1	0	2	1
<i>Dichanthelium boscii</i>	Poaceae	0	0	0	2	5
<i>Dichanthelium commutatum</i>	Poaceae	2	1	6	0	2
<i>Dichanthelium dichotomum</i>	Poaceae	0	0	1	0	0
<i>Dichanthelium ensifolium</i>	Poaceae	3	2	2	1	0
<i>Dichanthelium laxiflorum</i>	Poaceae	0	1	0	0	2
<i>Dichanthelium oligosanthes</i>	Poaceae	1	2	1	0	0
<i>Dichanthelium ovale</i>	Poaceae	5	3	0	0	0
<i>Dichanthelium ravenelii</i>	Poaceae	1	1	1	0	0
<i>Dichanthelium sphaerocarpon</i>	Poaceae	1	4	1	0	0
<i>Dichanthelium strigosum</i>	Poaceae	1	1	0	0	0
<i>Dichondra carolinensis</i>	Convolvulaceae	0	0	0	0	2
<i>Digitaria ciliaris</i>	Poaceae	0	0	0	0	1
<i>Digitaria filiformis</i>	Poaceae	2	4	0	0	0
<i>Diodia teres</i>	Rubiaceae	2	2	1	4	4

Appendix A. Continued.

Species	Family	Plots ^a			Transects ^a	
		N	O	U	N	O
<i>Diodia virginiana</i>	Rubiaceae	0	0	0	3	1
<i>Dioscorea bulbifera</i> *	Dioscoreaceae	0	0	1	0	0
<i>Dioscorea floridana</i>	Dioscoreaceae	0	0	0	1	0
<i>Diospyros virginiana</i>	Ebenaceae	8	8	8	6	9
<i>Drosera brevifolia</i>	Droseraceae	0	0	0	1	0
<i>Duchesnea indica</i> *	Rosaceae	0	0	0	1	0
<i>Dyschoriste oblongifolia</i>	Acanthaceae	11	0	2	5	2
<i>Elaeagnus pungens</i> *	Elaeagnaceae	0	0	0	0	1
<i>Elaeagnus umbellata</i> *	Elaeagnaceae	0	0	0	0	1
<i>Elephantopus carolinianus</i>	Asteraceae	0	0	0	0	1
<i>Elephantopus elatus</i>	Asteraceae	9	5	4	6	3
<i>Elephantopus nudatus</i>	Asteraceae	0	0	0	1	1
<i>Elephantopus tomentosus</i>	Asteraceae	0	0	0	1	1
<i>Eragrostis eliottii</i>	Poaceae	0	1	0	0	0
<i>Eragrostis virginica</i>	Poaceae	0	0	0	1	0
<i>Erythites hieracifolia</i>	Asteraceae	0	0	0	2	4
<i>Erigeron strigosus</i>	Asteraceae	0	0	0	0	2
<i>Eriocaulon decangulare</i>	Eriocaulaceae	0	0	0	1	0
<i>Eryngium yuccifolium</i>	Apiaceae	1	0	0	4	1
<i>Erythrina herbacea</i>	Fabaceae	1	7	1	3	6
<i>Eupatorium album</i>	Asteraceae	4	0	1	5	3
<i>Eupatorium altissimum</i>	Asteraceae	0	0	0	0	1
<i>Eupatorium capillifolium</i>	Asteraceae	1	7	1	1	1
<i>Eupatorium compostifolium</i>	Asteraceae	2	1	2	6	8
<i>Eupatorium hyssopifolium</i>	Asteraceae	4	8	5	4	7
<i>Eupatorium mohrii</i>	Asteraceae	0	0	0	3	1
<i>Eupatorium perfoliatum</i>	Asteraceae	0	0	0	0	1
<i>Eupatorium rotundifolium</i>	Asteraceae	4	2	1	4	2
<i>Eupatorium semiserratum</i>	Asteraceae	0	0	0	2	0
<i>Euphorbia curtisii</i>	Euphorbiaceae	0	0	0	3	0
<i>Euphorbia discoidalis</i>	Euphorbiaceae	6	0	2	0	0
<i>Euphorbia pubentissima</i>	Euphorbiaceae	0	0	0	4	1
<i>Euthamia caroliniana</i>	Asteraceae	1	0	0	0	1
<i>Fagus grandifolia</i>	Fagaceae	0	0	0	0	2
<i>Fraxinus americana</i>	Oleaceae	0	0	1	1	3
<i>Galactia erecta</i>	Fabaceae	1	0	1	1	0
<i>Galactia regularis</i>	Fabaceae	0	0	1	1	1
<i>Galactia villosa</i>	Fabaceae	3	1	2	3	6
<i>Galium pilosum</i>	Rubiaceae	11	7	10	8	7
<i>Gamochaeta americana</i>	Asteraceae	0	0	0	1	0
<i>Gamochaeta falcata</i>	Asteraceae	0	0	0	0	1
<i>Gamochaeta pennsylvanica</i>	Asteraceae	0	1	0	0	3
<i>Gaura angustifolia</i>	Onagraceae	0	0	0	1	0
<i>Gaura filipes</i>	Onagraceae	2	0	0	0	1
<i>Gaylussacia dumosa</i>	Ericaceae	5	0	0	3	0
<i>Gaylussacia frondosa</i>	Ericaceae	4	0	0	4	0
<i>Gelsemium sempervirens</i>	Gelsemiaceae	4	5	12	8	8
<i>Gentiana villosa</i>	Gentianaceae	0	0	0	0	1
<i>Geranium carolinianum</i>	Geraniaceae	0	0	0	0	1
<i>Gratiola pilosa</i>	Veronicaceae	0	0	0	1	0
<i>Gymnopogon ambiguus</i>	Poaceae	4	5	3	4	5
<i>Gymnopogon brevifolius</i>	Poaceae	3	0	0	0	0
<i>Hamamelis virginiana</i>	Hamamelidaceae	0	0	0	1	1
<i>Helianthemum carolinianum</i>	Cistaceae	1	0	0	0	0
<i>Helianthus angustifolius</i>	Asteraceae	3	0	1	2	0
<i>Helianthus radula</i>	Asteraceae	0	0	0	2	0
<i>Heteropogon melanocarpus</i>	Poaceae	0	0	0	1	0
<i>Hexalectris spicata</i>	Orchidaceae	0	0	0	1	0
<i>Hibiscus aculeatus</i>	Malvaceae	0	0	0	3	2
<i>Hieracium gronovii</i>	Asteraceae	6	3	3	4	3
<i>Houstonia procumbens</i>	Rubiaceae	11	5	5	4	3
<i>Hypericum crux-andreae</i>	Clusiaceae	3	2	1	3	0
<i>Hypericum galioides</i>	Clusiaceae	0	0	0	1	1
<i>Hypericum gentianoides</i>	Clusiaceae	0	0	0	2	1
<i>Hypericum hypericoides</i>	Clusiaceae	2	9	6	5	8
<i>Hypericum microsepalum</i>	Clusiaceae	0	0	0	0	1
<i>Hypericum mutilum</i>	Clusiaceae	0	0	0	0	1
<i>Hypericum setosum</i>	Clusiaceae	0	0	0	1	0
<i>Hypericum suffruticosum</i>	Clusiaceae	0	0	0	1	0

Appendix A. Continued.

Species	Family	Plots ^a			Transects ^a	
		N	O	U	N	O
<i>Hypericum tetrapetalum</i>	Clusiaceae	0	0	0	0	1
<i>Hyptis mutabilis</i> *	Lamiaceae	0	0	0	0	2
<i>Hypoxis juncea</i>	Hypoxidaceae	1	0	0	0	0
<i>Ilex cassine</i>	Aquifoliaceae	0	0	0	1	0
<i>Ilex coriacea</i>	Aquifoliaceae	0	0	0	2	1
<i>Ilex glabra</i>	Aquifoliaceae	5	0	1	5	2
<i>Ilex opaca</i>	Aquifoliaceae	0	0	0	1	5
<i>Ilex vomitoria</i>	Aquifoliaceae	0	0	0	0	2
<i>Indigofera caroliniana</i>	Fabaceae	0	0	0	0	1
<i>Ipomoea pandurata</i>	Convolvulaceae	0	0	0	0	1
<i>Ipomoea purpurea</i>	Convolvulaceae	0	0	0	1	0
<i>Itea virginica</i>	Iteaceae	0	0	0	2	1
<i>Jacquemontia tamnifolia</i>	Convolvulaceae	0	0	0	0	2
<i>Juncus coriaceous</i>	Juncaceae	0	0	0	0	2
<i>Juncus repens</i>	Juncaceae	0	0	0	1	0
<i>Juniperus virginiana</i>	Cupressaceae	0	0	0	0	1
<i>Kummerowia striata</i> *	Fabaceae	0	0	0	0	2
<i>Lactuca canadensis</i>	Asteraceae	0	1	0	0	1
<i>Lactuca floridana</i>	Asteraceae	0	1	0	0	0
<i>Lactuca graminifolia</i>	Asteraceae	0	0	0	1	2
<i>Lechea minor</i>	Cistaceae	2	0	0	1	0
<i>Lechea mucronata</i>	Cistaceae	0	0	0	2	4
<i>Lechea pulchella</i>	Cistaceae	1	0	0	0	0
<i>Leersia virginica</i>	Poaceae	0	0	0	0	1
<i>Lespedeza bicolor</i> *	Fabaceae	0	0	1	0	0
<i>Lespedeza capitata</i>	Fabaceae	0	0	0	1	0
<i>Lespedeza cuneata</i> *	Fabaceae	0	0	0	1	0
<i>Lespedeza hirta</i>	Fabaceae	4	3	4	0	2
<i>Lespedeza procumbens</i>	Fabaceae	0	0	1	0	3
<i>Lespedeza repens</i>	Fabaceae	1	1	0	1	2
<i>Lespedeza stuevei</i>	Fabaceae	0	0	1	0	0
<i>Lespedeza violacea</i>	Fabaceae	1	4	1	0	0
<i>Lespedeza virginica</i>	Fabaceae	2	2	4	0	1
<i>Leucothoe racemosa</i>	Ericaceae	0	0	0	1	0
<i>Liatris elegans</i>	Asteraceae	2	3	0	4	0
<i>Liatris gracilis</i>	Asteraceae	0	0	0	3	2
<i>Liatris graminifolia</i>	Asteraceae	0	0	1	2	2
<i>Liatris tenuifolia</i>	Asteraceae	2	0	0	3	0
<i>Ligustrum lucidum</i> *	Oleaceae	0	0	0	0	1
<i>Ligustrum sinense</i> *	Oleaceae	0	0	0	1	1
<i>Linum floridanum</i>	Linaceae	0	0	0	1	0
<i>Liquidambar styraciflua</i>	Altingiaceae	5	11	11	8	9
<i>Liriodendron tulipifera</i>	Magnoliaceae	0	0	0	1	2
<i>Lobelia amoena</i>	Campanulaceae	0	0	0	4	0
<i>Lobelia puberula</i>	Campanulaceae	1	1	1	1	0
<i>Lonicera japonica</i> *	Caprifoliaceae	0	0	0	1	2
<i>Lonicera sempervirens</i>	Caprifoliaceae	0	0	0	1	3
<i>Ludwigia hirtella</i>	Onagraceae	0	0	0	1	0
<i>Ludwigia pilosa</i>	Onagraceae	0	0	0	1	0
<i>Ludwigia virgata</i>	Onagraceae	0	0	0	1	0
<i>Lycopodiella alopecuroides</i>	Lycopodiaceae	1	0	0	0	0
<i>Lygodium japonicum</i> *	Schizaeaceae	0	0	1	4	5
<i>Lyonia lucida</i>	Ericaceae	0	0	0	4	1
<i>Magnolia grandiflora</i>	Magnoliaceae	0	0	0	4	3
<i>Magnolia virginiana</i>	Magnoliaceae	0	0	0	2	3
<i>Malus angustifolia</i>	Rosaceae	0	0	0	4	2
<i>Manfreda virginica</i>	Agavaceae	0	0	0	2	1
<i>Mecardonia acuminata</i>	Veronicaceae	0	0	0	1	0
<i>Melia azedarach</i> *	Meliaceae	0	0	0	0	1
<i>Mikania scandens</i>	Asteraceae	0	0	0	0	1
<i>Mimosa quadrivalvis</i>	Fabaceae	11	3	2	5	4
<i>Mitchella repens</i>	Rubiaceae	0	0	7	2	3
<i>Monarda punctata</i>	Lamiaceae	0	0	0	1	0
<i>Monotropa uniflora</i>	Ericaceae	3	0	0	2	0
<i>Morus rubra</i>	Moraceae	0	0	0	1	1
<i>Muhlenbergia capillaries</i>	Poaceae	4	0	1	3	0
<i>Myrica carolinensis</i>	Myricaceae	0	0	0	0	1
<i>Myrica cerifera</i>	Myricaceae	10	8	7	5	7
<i>Nyssa sylvatica</i>	Cornaceae	0	0	0	5	6

Appendix A. Continued.

Species	Family	Plots ^a			Transects ^a	
		N	O	U	N	O
<i>Oenothera biennis</i>	Onagraceae	0	0	0	1	2
<i>Oplismenus hirtellus</i>	Poaceae	0	0	0	0	1
<i>Osmanthus americanus</i>	Oleaceae	0	0	0	0	2
<i>Osmunda cinnamomea</i>	Osmundaceae	0	0	0	1	3
<i>Osmunda regalis</i>	Osmundaceae	0	0	0	0	1
<i>Ostrya virginiana</i>	Betulaceae	0	0	0	0	1
<i>Oxalis corniculata</i>	Oxalidaceae	0	2	1	1	4
<i>Oxydendrum arboretum</i>	Ericaceae	0	0	0	2	2
<i>Panicum anceps</i>	Poaceae	4	1	7	2	5
<i>Panicum tenerum</i>	Poaceae	1	0	0	0	0
<i>Panicum verrocosum</i>	Poaceae	0	1	0	0	1
<i>Panicum virgatum</i>	Poaceae	0	0	0	4	0
<i>Parthenocissus quinquefolia</i>	Vitaceae	1	6	5	2	8
<i>Paspalum bifidum</i> *	Poaceae	0	0	1	0	0
<i>Paspalum dilatatum</i> *	Poaceae	0	0	0	1	0
<i>Paspalum floridanum</i>	Poaceae	0	1	0	1	1
<i>Paspalum notatum</i> *	Poaceae	0	0	0	3	5
<i>Paspalum plicatulum</i>	Poaceae	0	0	0	0	1
<i>Paspalum praecox</i>	Poaceae	0	1	0	0	0
<i>Paspalum setaceum</i>	Poaceae	2	3	1	2	1
<i>Paspalum urvillei</i> *	Poaceae	0	0	0	1	1
<i>Passiflora incarnata</i>	Passifloraceae	0	1	0	1	5
<i>Passiflora lutea</i>	Passifloraceae	0	0	0	0	1
<i>Pediomelum canescens</i>	Fabaceae	0	0	0	2	0
<i>Penstemon multiflorus</i>	Veronicaceae	0	0	0	1	0
<i>Persea palustris</i>	Lauraceae	0	0	0	1	1
<i>Phanopyrum gymnocarpon</i>	Poaceae	0	0	0	0	1
<i>Phlox floridana</i>	Polemoniaceae	1	0	0	0	0
<i>Phlox pilosa</i>	Polemoniaceae	0	0	0	1	0
<i>Photinia pyrifolia</i>	Rosaceae	0	1	4	1	0
<i>Phyllanthus tenellus</i> *	Euphorbiaceae	0	0	0	0	2
<i>Phyllanthus urinaria</i> *	Euphorbiaceae	0	0	0	1	2
<i>Physalis arenicola</i>	Solanaceae	1	0	0	0	0
<i>Phytolacca americana</i>	Phytolaccaceae	0	0	0	1	2
<i>Pinus echinata</i>	Pinaceae	4	7	10	7	8
<i>Pinus elliotii</i>	Pinaceae	0	0	0	0	2
<i>Pinus glabra</i>	Pinaceae	0	0	0	1	1
<i>Pinus palustris</i>	Pinaceae	10	0	2	8	7
<i>Pinus taeda</i>	Pinaceae	5	4	2	6	9
<i>Piptochaetium avenaceum</i>	Poaceae	0	0	0	1	0
<i>Piriqueta cistoides</i>	Turneraceae	0	0	0	4	6
<i>Pityopsis aspera</i>	Asteraceae	11	7	5	4	2
<i>Pityopsis graminifolia</i>	Asteraceae	5	0	0	3	5
<i>Plantago virginica</i>	Plantaginaceae	0	1	0	0	0
<i>Platanthera ciliaris</i>	Orchidaceae	0	0	0	0	1
<i>Platanthera cristata</i>	Orchidaceae	0	0	0	1	0
<i>Pleopeltis polypodioides</i>	Polypodiaceae	0	0	0	3	0
<i>Pluchea foetida</i>	Asteraceae	0	0	0	1	0
<i>Polygala grandiflora</i>	Polygalaceae	3	0	0	1	4
<i>Polygala incarnata</i>	Polygalaceae	0	0	1	6	6
<i>Polygala nana</i>	Polygalaceae	2	0	3	3	2
<i>Polygala polygama</i>	Polygalaceae	4	2	0	0	2
<i>Polypremum procumbens</i>	Tetrachondraceae	1	4	1	3	4
<i>Polystichum acrostichoides</i>	Dryopteridaceae	0	0	0	0	2
<i>Prenanthes serpentaria</i>	Asteraceae	0	0	0	0	1
<i>Prunus angustifolia</i>	Rosaceae	0	0	0	0	1
<i>Prunus caroliniana</i>	Rosaceae	0	0	0	0	1
<i>Prunus serotina</i>	Rosaceae	8	11	11	8	9
<i>Prunus umbellata</i>	Rosaceae	0	1	0	2	2
<i>Pseudognaphalium obtusifolium</i>	Asteraceae	0	0	0	0	1
<i>Pteridium aquilinum</i>	Dennstaedtiaceae	10	0	4	7	3
<i>Pycnanthemum flexuosum</i>	Lamiaceae	0	0	0	0	1
<i>Pycnanthemum floridanum</i>	Lamiaceae	0	0	0	1	0
<i>Pyrhopappus carolinianus</i>	Asteraceae	0	0	0	0	1
<i>Quercus alba</i>	Fagaceae	0	2	4	4	3
<i>Quercus elliotii</i>	Fagaceae	5	0	1	3	1
<i>Quercus falcata</i>	Fagaceae	8	7	7	5	6

Appendix A. Continued.

Species	Family	Plots ^a			Transects ^a	
		N	O	U	N	O
<i>Quercus incana</i>	Fagaceae	10	1	2	4	0
<i>Quercus laevis</i>	Fagaceae	5	0	0	4	0
<i>Quercus laurifolia</i>	Fagaceae	4	3	8	4	5
<i>Quercus margaretta</i>	Fagaceae	4	3	1	2	1
<i>Quercus marilandica</i>	Fagaceae	1	0	0	0	0
<i>Quercus michauxii</i>	Fagaceae	0	0	0	0	1
<i>Quercus nigra</i>	Fagaceae	3	7	4	4	9
<i>Quercus stellata</i>	Fagaceae	11	7	8	5	3
<i>Quercus velutina</i>	Fagaceae	0	0	0	1	1
<i>Quercus virginiana</i>	Fagaceae	0	6	5	4	6
<i>Rhapidophyllum hystrix</i>	Arecaceae	0	0	0	0	1
<i>Rhexia virginica</i>	Melastomataceae	0	0	0	1	0
<i>Rhododendron canescens</i>	Ericaceae	0	0	0	1	0
<i>Rhus copallinum</i>	Anacardiaceae	12	11	12	7	7
<i>Rhus glabra</i>	Anacardiaceae	0	0	0	1	1
<i>Rhynchosia difformis</i>	Fabaceae	1	0	0	2	1
<i>Rhynchosia reniformis</i>	Fabaceae	11	1	2	4	4
<i>Rhynchosia tomentosa</i>	Fabaceae	1	0	0	3	3
<i>Rhynchospora chalarocephala</i>	Cyperaceae	0	0	0	1	0
<i>Rhynchospora grayi</i>	Cyperaceae	2	0	0	0	1
<i>Rhynchospora harveyi</i>	Cyperaceae	0	0	0	1	0
<i>Rhynchospora intermedia</i>	Cyperaceae	1	0	0	0	0
<i>Rhynchospora miliacea</i>	Cyperaceae	0	0	0	0	2
<i>Rhynchospora plumosa</i>	Cyperaceae	1	0	0	0	0
<i>Richardia scabra</i> *	Rubiaceae	0	0	0	1	1
<i>Rubus argutus</i>	Rosaceae	7	8	9	1	3
<i>Rubus cuneifolius</i>	Rosaceae	8	11	11	8	9
<i>Rubus flagellaris</i>	Rosaceae	0	0	0	4	4
<i>Rubus trivialis</i>	Rosaceae	1	0	0	1	5
<i>Rudbeckia hirta</i>	Asteraceae	1	0	0	4	1
<i>Ruellia caroliniensis</i>	Acanthaceae	1	0	0	0	3
<i>Ruellia ciliosa</i>	Acanthaceae	1	0	0	0	0
<i>Rumex hastatulus</i>	Polygonaceae	0	0	0	0	1
<i>Sabal minor</i>	Arecaceae	0	0	0	1	2
<i>Sabatia angularis</i>	Gentianaceae	0	0	0	0	2
<i>Sabatia calycina</i>	Gentianaceae	0	0	0	0	1
<i>Saccharum alopecuroides</i>	Poaceae	3	10	4	3	5
<i>Saccharum brevibarbe</i>	Poaceae	0	0	0	1	0
<i>Saccharum coarctatum</i>	Poaceae	3	0	0	0	0
<i>Salix humilis</i>	Salicaceae	0	0	0	2	1
<i>Salvia azurea</i>	Lamiaceae	6	0	1	6	0
<i>Salvia lyrata</i>	Lamiaceae	0	0	0	0	2
<i>Sambucus nigra</i>	Adoxaceae	0	0	0	0	2
<i>Sanicula canadensis</i>	Apiaceae	0	0	0	1	2
<i>Sanicula smallii</i>	Apiaceae	0	0	0	0	1
<i>Sassafras albidum</i>	Lauraceae	4	5	1	6	7
<i>Schizachyrium sanguineum</i>	Poaceae	0	0	0	1	0
<i>Schizachyrium scoparium</i>	Poaceae	9	5	3	0	0
<i>Schizachyrium tenerum</i>	Poaceae	2	0	1	1	0
<i>Scirpus cyperinus</i>	Cyperaceae	0	0	0	1	0
<i>Scleria ciliata</i>	Cyperaceae	4	4	0	0	0
<i>Scleria oligantha</i>	Cyperaceae	0	2	0	0	0
<i>Scleria reticularis</i>	Cyperaceae	0	0	0	1	0
<i>Scleria triglomerata</i>	Cyperaceae	0	0	0	1	3
<i>Scutellaria integrifolia</i>	Lamiaceae	0	0	0	2	3
<i>Scutellaria multiglandulosa</i>	Lamiaceae	0	0	0	1	0
<i>Sebastiania fruticosa</i>	Euphorbiaceae	0	0	0	3	0
<i>Senna obtusifolia</i> *	Fabaceae	0	0	0	1	2
<i>Senna occidentalis</i> *	Fabaceae	0	1	0	0	0
<i>Serenoa repens</i>	Arecaceae	0	0	0	1	0
<i>Sericocarpus tortifolius</i>	Asteraceae	10	0	7	7	5
<i>Setaria parviflora</i>	Poaceae	1	0	0	0	0
<i>Setaria pumila</i> *	Poaceae	0	0	0	0	2
<i>Seymeria cassioides</i>	Orobanchaceae	0	0	0	1	0
<i>Seymeria pectinata</i>	Orobanchaceae	2	0	0	3	0
<i>Sideroxylon lanuginosum</i>	Sapotaceae	0	0	1	0	1
<i>Silphium asteriscus</i>	Asteraceae	0	0	0	4	1
<i>Smilax auriculata</i>	Smilacaceae	3	1	3	7	5
<i>Smilax bona-nox</i>	Smilacaceae	8	5	4	7	5

Appendix A. Continued.

Species	Family	Plots ^a			Transects ^a	
		N	O	U	N	O
<i>Smilax glauca</i>	Smilacaceae	4	8	8	8	8
<i>Smilax laurifolia</i>	Smilacaceae	0	2	0	2	0
<i>Smilax pumila</i>	Smilacaceae	0	1	1	3	3
<i>Smilax rotundifolia</i>	Smilacaceae	0	0	0	3	1
<i>Smilax smallii</i>	Smilacaceae	2	5	3	5	7
<i>Smilax tamnoides</i>	Smilacaceae	5	4	7	3	4
<i>Smilax walteri</i>	Smilacaceae	0	0	0	1	0
<i>Solanum carolinense</i>	Solanaceae	0	0	0	1	3
<i>Solidago arguta</i>	Asteraceae	0	0	0	3	3
<i>Solidago auriculata</i>	Asteraceae	0	0	0	0	1
<i>Solidago canadensis</i>	Asteraceae	2	4	3	1	7
<i>Solidago odora</i>	Asteraceae	8	7	7	7	8
<i>Solidago rugosa</i>	Asteraceae	0	5	3	1	4
<i>Solidago stricta</i>	Asteraceae	1	1	1	1	0
<i>Solidago tortifolia</i>	Asteraceae	1	0	0	0	1
<i>Sonchus oleraceus</i>	Asteraceae	0	0	0	0	1
<i>Sorghastrum elliotii</i>	Poaceae	0	0	0	2	1
<i>Sorghastrum nutans</i>	Poaceae	2	1	7	0	0
<i>Sorghastrum secundum</i>	Poaceae	2	7	7	0	0
<i>Sorghum halepense*</i>	Poaceae	0	0	0	1	1
<i>Spiranthes tuberosa</i>	Orchidaceae	0	0	0	0	2
<i>Sporobolus floridanus</i>	Poaceae	0	0	0	1	0
<i>Sporobolus indicus*</i>	Poaceae	0	0	0	1	0
<i>Sporobolus junceus</i>	Poaceae	1	0	0	1	0
<i>Stillingia sylvatica</i>	Euphorbiaceae	2	0	0	4	1
<i>Strophostyles umbellata</i>	Fabaceae	9	4	2	1	0
<i>Stylisma humistrata</i>	Convolvulaceae	0	0	0	1	0
<i>Stylisma patens</i>	Convolvulaceae	3	0	0	1	0
<i>Stylosanthes biflora</i>	Fabaceae	7	1	6	5	4
<i>Symphyotrichum adnatum</i>	Asteraceae	9	0	0	4	2
<i>Symphyotrichum concolor</i>	Asteraceae	11	1	3	5	2
<i>Symphyotrichum dumosum</i>	Asteraceae	5	4	8	5	1
<i>Symphyotrichum patens</i>	Asteraceae	0	0	0	1	1
<i>Symphyotrichum sagittifolium</i>	Asteraceae	0	0	0	4	3
<i>Symphyotrichum undulatum</i>	Asteraceae	0	0	0	1	0
<i>Symplocos tinctoria</i>	Symplocaceae	1	1	3	2	2
<i>Taxodium ascendens</i>	Cupressaceae	0	0	0	1	0
<i>Tephrosia hispidula</i>	Fabaceae	1	0	0	0	2
<i>Tephrosia spicata</i>	Fabaceae	2	1	0	2	2
<i>Tephrosia virginiana</i>	Fabaceae	4	0	0	5	2
<i>Tetragonotheca helianthoides</i>	Asteraceae	0	0	0	0	1
<i>Teucrium canadense</i>	Lamiaceae	1	0	0	0	0
<i>Thelypteris kunthii</i>	Thelypteridaceae	0	0	0	0	1
<i>Tillandsia usneoides</i>	Bromeliaceae	0	0	0	5	3
<i>Toxicodendron pubescens</i>	Anacardiaceae	0	0	0	2	1
<i>Toxicodendron radicans</i>	Anacardiaceae	9	6	8	2	5
<i>Toxicodendron vernix</i>	Anacardiaceae	0	0	0	0	1
<i>Tragia smallii</i>	Euphorbiaceae	0	0	0	0	1
<i>Tragia urens</i>	Euphorbiaceae	4	2	1	2	0
<i>Tragia urticifolia</i>	Euphorbiaceae	0	0	0	0	2
<i>Trichostema dichotomum</i>	Lamiaceae	0	6	1	3	3
<i>Trichostema setaceum</i>	Lamiaceae	2	0	0	0	0
<i>Tridens carolinianus</i>	Poaceae	1	0	0	0	0
<i>Tridens flavus</i>	Poaceae	3	7	6	4	2
<i>Utricularia purpurea</i>	Lentibulariaceae	0	0	0	1	0
<i>Vaccinium arboreum</i>	Ericaceae	5	1	6	8	6
<i>Vaccinium corymbosum</i>	Ericaceae	3	0	0	2	1
<i>Vaccinium darrowii</i>	Ericaceae	5	0	4	2	1
<i>Vaccinium myrsinites</i>	Ericaceae	5	1	0	6	2
<i>Vaccinium stamineum</i>	Ericaceae	0	2	6	6	8
<i>Verbesina aristata</i>	Asteraceae	4	0	1	4	4
<i>Verbesina virginica</i>	Asteraceae	0	0	0	1	2
<i>Vernonia angustifolia</i>	Asteraceae	4	1	2	6	6
<i>Viburnum dentatum</i>	Adoxaceae	0	0	0	0	1
<i>Viburnum nudum</i>	Adoxaceae	0	0	0	0	1
<i>Viburnum rufidulum</i>	Adoxaceae	0	0	0	2	3
<i>Viola palmata</i>	Violaceae	1	0	1	2	0
<i>Viola primulifolia</i>	Violaceae	0	0	0	0	2
<i>Viola sororia</i>	Violaceae	0	0	0	0	1

Appendix A. Continued.

Species	Family	Plots ^a			Transects ^a	
		N	O	U	N	O
<i>Viola villosa</i>	Violaceae	0	0	0	2	1
<i>Vitis aestivalis</i>	Vitaceae	0	3	1	1	3
<i>Vitis cinerea</i>	Vitaceae	0	0	0	0	1
<i>Vitis rotundifolia</i>	Vitaceae	2	5	9	7	10
<i>Wahlenbergia marginata</i>	Campanulaceae	0	0	0	1	3
<i>Woodwardia areolata</i>	Blechnaceae	0	0	0	0	1
<i>Woodwardia virginica</i>	Blechnaceae	0	0	0	1	1
<i>Xyris caroliniana</i>	Xyridaceae	0	0	0	1	0
<i>Yucca filamentosa</i>	Agavaceae	0	0	0	0	3

^a Abbreviations: N, native ground cover; O, old-field ground cover; U, unknown soil disturbance history.